# Estimation of mercury transport and accumulation from an incinerator

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#### ABSTRACT

Mercury concentrations in the air, soil, and sediment - water were estimated to determine the levels of mercury that might be present in the environmental media within 5 km of an incinerator in New Jersey. Gaussian dispersion equation was used as a basic tool for estimating mercury concentrations in the air Universal loss equation, atmospheric deposition (wet and dry) and accumulation equations were used to calculate mercury concentrations in soil.

and sediment - water

Estimated high-ground level concentration of mercury in the air was about 28 ng/m³ under D stability, which is slightly higher than the average in the U.S. (2 - 20 ng/m³) due to the result of near - source calculations (5 km). Mercury in soil and sediment, regardless background concentration, were approximately 0.37 and 0.33 mg/kg, respectively, which is high comparing to another county soil samples

#### INTRODUCTION

Mercury is considered one of the toxic metals and it can accumulate in the food chain (especially methymercury). In human, mercury can cause paresthesia and developmental neurologic effects in - utero (2) Serveral studied showed that mercury emitted from incinerators causes the increasing of mercury contamination in the environment (3-4) In Minamata Bay, Japan, widespread problem of

mercury contamination caused Minamata disease Mercury pollution has become the subject of much concern and public outcry. This paper attempted to estimate mercury concentrations in air, soil, and sediment - water within 5 km downwind from and incinerator in New Jersey with the best available data from literatures.

#### **METHODS**

#### 1. Determination of mercury concentrations

#### in the air

Gaussian dispersion equation (1) (5,6) was used to estimate the mercury levels downwind. The calculations assumed that the plume from the incin-

erator behaved similarly to that described by the Gaussian plume model (Figure 1) for simple terrain and that no chemical reactions occur during the transport (5km):

$$C = \frac{Q}{2\pi u_s \sigma_y \sigma_z} e^{\left[-\frac{1}{2} \left(\frac{y}{\sigma_y}\right)^2\right]} \cdot \left[ e^{\left[-\frac{1}{2} \left(\frac{z-H}{\sigma_z}\right)^2\right] + R \cdot e^{\left[-\frac{1}{2} \left(\frac{z+H}{\sigma_z}\right)^2\right]}\right] - - (1)$$

where C is the downwind concentration of total mercury  $(g/m^3)$ , Q is the total mercury emission rate (g/s),  $U_s$  is the average wind speed at stack height (m/s),  $\mathbf{O}_y$  is the horizontal distance from plume centerline (m), Z is the vertical distance from plume centerline (m), R is the reflection of plume when it

reaches the ground (assumed 10% absorption due to soluble species,  $Hg^{2}$ ), and H is the effective stack height (m). Equation (2) and (3) were derived from the User's Guide for ISC3 (1995)<sup>(7)</sup> to estimate  $\sigma$ , and  $\sigma$ :

where x is the distance in km; a, b, c, and d are coeffcients depending on downwind distance and stability, which are used to fit Pasquill - Gifford curves. (7)

In addition effective stack height was calculated from the following equation :  $^{(8)}\,$ 

$$H = \Delta H v + h \qquad --- (4)$$

where h is the stack height (m), and  $\Delta H$  is the plume rise (m) calculated by :

$$\Delta H = \frac{v_s d}{u} \left[ 1.5 + \left( 2.68 \times 10^{-2} (P) \left( \frac{T_s - T_a}{T_s} \right) d \right) \right] - (5)$$

mercury in soil and sediment - water

where  $\Delta H$  is the plume rise (m),  $v_s$  is the stack exit velocity (m/s), d is the stack diameter (m), P is the atmospheric pressure (kPa).  $T_s$  is the stack temperature (\*K), and  $T_s$  is the atmospheric temperature (\*K). Table 1 shows the parameters used in the calculations.

#### 2. Determination of mercury depositions

Once mercury is in the air, it will eventually deposit to the ground, water bodies, plants, and so on The wet and dry deposition rate of atmospheric mercury were applied to estimate the accumulation of

a) Wet deposition: Soluble forms of mercury such as  $\mathrm{HgCl}_2$ , may be removed by precipitation, whereas insoluble, volatile forms such as elemental mercury ( $\mathrm{Hg}^*$ ) and dimethylmercury [( $\mathrm{CH}_2$ )<sub>2</sub> $\mathrm{Hg}$ ] are likely to remain in the atmosphere <sup>(9)</sup> Wet removal of mercury can be estimated using a washout ratio (WR) approach WR is defined as  $\mathrm{C}_{\mathrm{rainwater}}$  [ $\mathrm{ng/m^3}$ ]/ $\mathrm{C}_{\mathrm{air}}$  [ $\mathrm{ng/m^3}$ ] ratio, <sup>(10)</sup> 111 The wet deposition flux (WD) was then estimated from the following equation. <sup>(CO)</sup>

$$WD = (C_{al})(WR)(R)(t) \qquad --- (6)$$

where  $C_{_{all}}$  is the murcury concentration in air (ng/m³), R is the annual average rainfall rate (m/hr), and t is the time accounted for rainfall (yr i)

b) Dry deposition Atmospheric mercury

accounts only for a small portion of particulate Hg° (<5%) compared to the gaseous phase (\*2) Dry deposition flux (DD) was estimated based on deposition velocity (V) as described by NJDEPE (1993) (19):

Dry deposition flux (DD) = 
$$(C_{ad})(V_{ad})$$
 --- (7)

Total mercury deposition (TD) is the total of WD and DD

## 3. Determination of mercury concentrations in soil (due to depositions)

Universal Loss Equation Developed by

 $USDA^{c.c.}$  was employed to estimate mercury concentration in soil

$$\chi_{c} = (E_{R}) (K_{SE}) (LS) (c) (P_{S}) (D_{DE}) \qquad --- (8)$$

where  $\chi_{\rm e}$  is the loss rate per unit area watershed over time (g/m²-yr),  $E_{\rm i}$  is the rainfall erosion index (yr¹), (10)  $K_{\rm SE}$  is the soil erodability factor (ton/acre) based on type of soil and organic content, c is the unitless cover and management factor based on

the type of land use and the percentage of ground cover,  $P_{\underline{\iota}}$  is the unitless supporting practice factor,  $D_{\underline{DF}}$  is the unitless sediment delivery factor, LS is the slope length factor .

LS = 
$$(\lambda/22.1)^{\xi}$$
 (65.41S<sup>2</sup> + 4.565S + 0.065) --- (9)

where LS is the unitless slope length factor to determine the rate of soil loss  $\chi_{\rm e}$ ,  $\lambda$  is the land surface slope length (m),  $\xi$  is the exponent to

determine slope length (LS), and S is the land surface slope (m/m).

$$\xi$$
 = 0.6 [1-e<sup>(-35 835S)</sup>] --- (10)

The first - order loss rate of mercury calculated erosion  $(k_1 - yr^4)$  was based on the following equation :<sup>(10)</sup> in soil due to soil

$$k_1 = (\chi_{\alpha})/[(BD_{\alpha})] (SD_{\alpha})] --- (11)$$

where  $BD_{soil}$  is the soil mixing depth (m), and concentration from total deposition was calculated . (10)  $SD_{soil}$  is the soil bulk density (kg/m). Mercury

$$C_{soil} = \frac{TD\left[\frac{1-e^{(-k_i)(AT)}}{k_i}\right]}{(SD_{soil})(BD_{soil})} --- (12)$$

where  $C_{soil}$  is the annual concentration of mercury in soil due to wet and dry deposition (mg/kg), TD is the total deposition rate of murcury to soil (g/m²-yr), and AT is the accumulation rate (yr).

4. Determination of mercury concentrations

#### in sediment - water

 $\label{eq:mercury} \mbox{Mercury concentrations in sediment - water} \\ \mbox{were calculated from the following:}$ 

a) Mercury concentrations in sediment - water due to runoff (C  $_{\rm col}$  :  $^{(10)}$ 

$$C_{sed} = \frac{(\chi_e)(WA_L)(C_{soil})(AT)}{(BD_{soil})(SD_{soil})(WA_w)} \qquad -- (13)$$

where the  $WA_L$  is watershed area contributing to runoff (m<sup>2</sup>), AT is the accumulation time (yr), and  $WA_W$  is the area of waterbody receiving runoff (m<sup>2</sup>).

$$C_{sed (dep)} = \frac{(\chi_e)(AT)(TD)}{(BD_{soil})(SD_{soil})} -- (14)$$

#### **RESULTS**

33

Figure 2 and 3 show the patterns and concentrations of mercury along the plume centerline for X - Z (100m x 50m scale), and X - Y (100m x 10m scale), respectively. The concentrations were calculated in every cell along both X - Z and X - Y axis. These calculations (ng/m $^3$ ) were examples of the concentrations under neutral stability (D - stability) and  $u_s$  of 2 m/s. The plume spread slowly under neutral stability, thus GLC further away (downwind) from the incinerator at 3.7 to 5 km were the highest (approximately 28 ng/m $^3$ ).

WR was calculated from data of mercury concentrations in the air and rain collected by Greengerg et al. (1992) (14) Estimated WR was  $5.04 \times 10^3$ . Then, WD was approximately  $4.2 \times 10^4$  g/m²/yr. The estimated DD using  $v_a$  of 0.006 cm/s was  $5.29 \times 10^4$  g/m²-yr. Thus, TD of atmospheric mercury to surfaces in the area was  $4.73 \times 10^{14}$  g/m²-yr.

Estimated LS was 339 that was used to estimate  $\chi_{\rm a}$  of 27.8 g/m'-yr BD  $_{\rm so}$  is 0.01 m for standard assumption and deposited mercury is obviously retained at the soil surface,  $^{(15-16)}$  SD  $_{\rm soil}$  is 1300 Km/m $^3$ for the most likely value  $^{\mbox{\scriptsize (10)}}$  Results showed the K was 2.14 x 10  $^{\rm f}$  yr and  $\rm C_{_{SOI}}$  was 0.37 mg/kg  $\rm \, K_{_{\rm SE}}$  was 0.35 ton/acre (assuming loam or silty loam). (13) The cvalue of 0.003, the most likely value, was used because the area around the incinerator was assumed to consist of 20 to 70% undisturbed land with greater than 40% litter cover. The default for P is 0.5<sup>(13)</sup> The D value of 0.4 was given by Bonazountas and Wagner (1984) "." Estimated  $C_{\text{sec}}$  was 0.27 mg/kg and  $C_{\text{equiv.}}$ was 0.063 mg/kg when WA of 10,000 m $^3$  and WA $_{\rm w}$  of 50,000 m° were used (derived from the GIS database of the surrounding watershed) 1101 Thus, total mercury in sediment water was approximatedly 0 33 mg/kg.

#### DISCUSSIONS AND CONCLUSIONS

Estimated (high) mercury concentrations of 28 ng/m³ (regardless background concentrations) were higher than average in the U.S. (2 - 20 ng/m³)(18) due to the near - source calculations and assumptions used in the calculations (no chemical reactions (loss), total mercury estimations). Background concentrations were reported by NJDEPE (1993) to be about 2 - 5 mg/m³. WD of mercury accounted for 89% of TD because the wet deposition processes are an important removal process of soluble - mercury species (such as HgCl) near the incinerator In addition, C, and C data from Greenberg *et al.* (1992) provided WR of 5.04 x 10³ which is higher than data reported value

(not in New Jersey) by Brosset (1987)<sup>(19)</sup> of 1 x 10<sup>4</sup> to 5 x 10<sup>4</sup> because the WR was derived from near - source measurement (2 to 3 km). From the results of the depositions, mercury in soil, and sediment - water of 0.37 to 0.33 mg/kg were also high comparing to another county (Atlantic County) soil samples (<0.062 to <0.089 mg/kg)<sup>(18)</sup> Results from the calculations indicated that high levels of mercury can be found in the air, soil, and sediment - water in the vicinity of the incinerator Thus, the incinerator can be considered a major source of mercury pollution in the area in accordant with the reports by Aucott and Winka (1996), and NJDEPE (1993) Further information about

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mercury in biota, e.g., fish, can be calculated using bioconcentration factor available in literatures.

Application of Gaussian dispersion equation shows that is can be a very useful tool for preliminary estimation of mercury dispersion point sources for certain assumptions. More accurate calculations can be accomplished with complex models that account for chemical reactions of mercury or complex terrain. Universal loss equation, depositions, and other accumulation equations are some of the methods used by regulatory agencies to estimate the concentrations of mercury in the environment

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### APPENDIX

Table 1. : Parameters used in calculations

Parameters	Values	Remarks
Q	0.013 µg/s	Premit Value (NJDEPE, 1993)
h	76.2 m	Radian Corporation (1989)
V <sub>s</sub>	18.29 m/s	Martin (1988)
d	2.48 m	Radian Corporation (1989)
R	0.9	Reflection of plume 90% (assume 10% soluble)
P	9.87 x 10 <sup>-3</sup> kPa	Standard pressure (1 atm)
$T_{_{\mathrm{s}}}$	388.6°K	Martin (1988)
$T_{a}$	293°K	Standard temperature (20°C)
u <sub>s</sub>	2 m/s	Chu (1996)
WR (	5.04 x 10 <sup>3</sup>	Greenberg et al , 1992
R <sub>f</sub>	1000 hr/yr	Chu (1996) NJ data
(F)	0.003 m/hr	Chu (1996) NJ data
v <sub>d</sub>	0.006 cm/s	Lindberg etaal., 1992
S	0.0433 m/m	Most likely value (NJDEPE, 1993)
	2000 m	NJDEPT, (1993)
E	175 yr <sup>-1</sup>	NJDEPE, (1993)
K <sub>SE</sub>	0.35 ton/acre	NJDEPE, (1993)
С	0.003	Most likely value (NJDEPE, 1993)
Р	0.5	NJDEPE, (1993)
D	0.4	Bonazauntas and Wagner, 1984
$\mathrm{BD}_{\mathrm{soil}}$	0.01 m	Standard value (NJDEPE, 1993)
$\mathrm{SD}_{\mathrm{soil}}$	1300 kg/m³	Most likely value (NJDEPE, 1993)
AT	10 yr	NJDEPE, (1993)
$WA_{_L}$	100,000 m <sup>2</sup>	NJDEPE, (1993)
WA <sub>w</sub>	50,000 m <sup>2</sup>	NJDEPE, (1993)

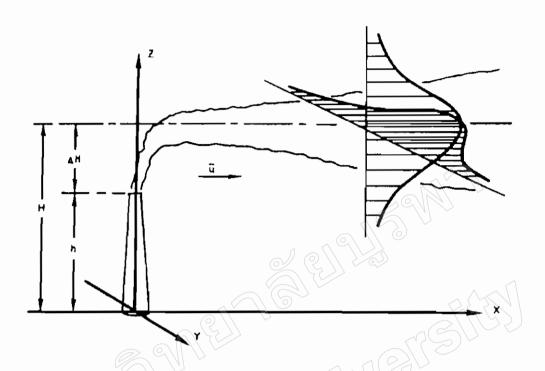


Figure 1: Gaussian Plume Model

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Estimated total mercury concentrations along x-y axis

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