

Estimation of mercury transport and accumulation from an incinerator

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ABSTRACT

Mercury concentrations in the air, soil, and sediment - water were estimated to determine the levels of mercury that might be present in the environmental media within 5 km of an incinerator in New Jersey. Gaussian dispersion equation was used as a basic tool for estimating mercury concentrations in the air. Universal loss equation, atmospheric deposition (wet and dry) and accumulation equations were used to calculate mercury concentrations in soil.

and sediment - water

Estimated high - ground level concentration of mercury in the air was about 28 ng/m³ under D - stability, which is slightly higher than the average in the U.S. (2 - 20 ng/m³) due to the result of near - source calculations (5 km). Mercury in soil and sediment, regardless background concentration, were approximately 0.37 and 0.33 mg/kg, respectively, which is high comparing to another county soil samples

INTRODUCTION

Mercury is considered one of the toxic metals and it can accumulate in the food chain (especially methylmercury).⁽¹⁾ In human, mercury can cause paresthesia and developmental neurologic effects in - utero.⁽²⁾ Several studies showed that mercury emitted from incinerators causes the increasing of mercury contamination in the environment.^(3, 4) In Minamata Bay, Japan, widespread problem of

mercury contamination caused Minamata disease. Mercury pollution has become the subject of much concern and public outcry. This paper attempted to estimate mercury concentrations in air, soil, and sediment - water within 5 km downwind from an incinerator in New Jersey with the best available data from literatures.

METHODS

1. Determination of mercury concentrations

in the air

Gaussian dispersion equation (1) ^(6,6) was used to estimate the mercury levels downwind. The calculations assumed that the plume from the incin-

erator behaved similarly to that described by the Gaussian plume model (Figure 1) for simple terrain and that no chemical reactions occur during the transport (5km) :

$$C = \frac{Q}{2\pi u_s \sigma_y \sigma_z} e^{\left[-\frac{1}{2} \left(\frac{y}{\sigma_y}\right)^2\right]} \cdot \left[e^{\left[-\frac{1}{2} \left(\frac{z-H}{\sigma_z}\right)^2\right]} + R \cdot e^{\left[-\frac{1}{2} \left(\frac{z+H}{\sigma_z}\right)^2\right]} \right] \quad \text{--- (1)}$$

where C is the downwind concentration of total mercury (g/m^3), Q is the total mercury emission rate (g/s), U_s is the average wind speed at stack height (m/s), σ_y is the horizontal distance from plume centerline (m), Z is the vertical distance from plume centerline (m), R is the reflection of plume when it

reaches the ground (assumed 10% absorption due to soluble species, Hg^{2+}), and H is the effective stack height (m). Equation (2) and (3) were derived from the User's Guide for ISC3 (1995)⁽⁷⁾ to estimate σ_y and σ_z :

$$\sigma_y = 465 \ 11628 \ (x) \ \text{TAN} \ (\text{TH}); \ \text{TH} = 0.017453293 \ [c - d \ \ln(x)] \quad \text{--- (2)}$$

$$\sigma_z = a(x)^b \quad \text{--- (3)}$$

where x is the distance in km; a, b, c, and d are coefficients depending on downwind distance and stability, which are used to fit Pasquill - Gifford curves.⁽⁷⁾

In addition effective stack height was calculated from the following equation : ⁽⁸⁾

$$H = \Delta H v + h \quad \text{--- (4)}$$

where h is the stack height (m), and ΔH is the plume rise (m) calculated by :

$$\Delta H = \frac{v_s d}{u} \left[1.5 + \left(2.68 \times 10^{-2} (P) \left(\frac{T_s - T_a}{T_s} \right) d \right) \right] \quad \text{--- (5)}$$

where ΔH is the plume rise (m), v_s is the stack exit velocity (m/s), d is the stack diameter (m), P is the atmospheric pressure (kPa), T_s is the stack temperature ($^{\circ}K$), and T_a is the atmospheric temperature ($^{\circ}K$). Table 1 shows the parameters used in the calculations.

2. Determination of mercury depositions

Once mercury is in the air, it will eventually deposit to the ground, water bodies, plants, and so on. The wet and dry deposition rate of atmospheric mercury were applied to estimate the accumulation of

mercury in soil and sediment - water

a) *Wet deposition* : Soluble forms of mercury such as $HgCl_2$, may be removed by precipitation, whereas insoluble, volatile forms such as elemental mercury (Hg^0) and dimethylmercury [$(CH_3)_2Hg$] are likely to remain in the atmosphere⁽⁴⁾. Wet removal of mercury can be estimated using a washout ratio (WR) approach. WR is defined as $C_{rain/water}$ [ng/m^3]/ C_{air} [ng/m^3] ratio,⁽¹⁰⁻¹¹⁾ The wet deposition flux (WD) was then estimated from the following equation⁽¹⁰⁾:

$$WD = (C_{air})(WR)(R)(t) \quad \text{--- (6)}$$

where C_{air} is the mercury concentration in air (ng/m^3), R is the annual average rainfall rate (m/hr), and t is the time accounted for rainfall (yr^{-1})

b) *Dry deposition* Atmospheric mercury

accounts only for a small portion of particulate Hg^0 (<5%) compared to the gaseous phase⁽¹²⁾. Dry deposition flux (DD) was estimated based on deposition velocity (V_d) as described by NJDEPE (1993)⁽¹³⁾:

$$\text{Dry deposition flux (DD)} = (C_{air})(V_d) \quad \text{--- (7)}$$

Total mercury deposition (TD) is the total of WD and DD

3. Determination of mercury concentrations in soil (due to depositions)

Universal Loss Equation Developed by

$$\chi_o = (E_r)(K_{SE})(LS)(c)(P_s)(D_{DF}) \quad \text{--- (8)}$$

where χ_o is the loss rate per unit area watershed over time ($g/m^2\text{-yr}$), E_r is the rainfall erosion index (yr^{-1}),⁽¹⁰⁾ K_{SE} is the soil erodability factor (ton/acre) based on type of soil and organic content, c is the unitless cover and management factor based on

USDA⁽¹⁴⁾ was employed to estimate mercury concentration in soil

the type of land use and the percentage of ground cover, P_s is the unitless supporting practice factor, D_{DF} is the unitless sediment delivery factor, LS is the slope length factor .

$$LS = (\lambda/22.1)^5 (65.41S^2 + 4.565S + 0.065) \quad \text{--- (9)}$$

where LS is the unitless slope length factor to determine the rate of soil loss χ_e , λ is the land surface slope length (m), ξ is the exponent to

determine slope length (LS), and S is the land surface slope (m/m).

$$\xi = 0.6 [1 - e^{(-35.835S)}] \quad \text{--- (10)}$$

The first - order loss rate of mercury calculated based on the following equation ⁽¹⁰⁾ in soil due to soil erosion ($k_1 - \text{yr}^{-1}$) was

$$k_1 = (\chi_e) / [(BD_{soil}) (SD_{soil})] \quad \text{--- (11)}$$

where BD_{soil} is the soil mixing depth (m), and SD_{soil} is the soil bulk density (kg/m³). Mercury concentration from total deposition was calculated ⁽¹⁰⁾

$$C_{soil} = \frac{TD \left[\frac{1 - e^{(-k_1)(AT)}}{k_1} \right]}{(SD_{soil}) (BD_{soil})} \quad \text{--- (12)}$$

where C_{soil} is the annual concentration of mercury in soil due to wet and dry deposition (mg/kg), TD is the total deposition rate of mercury to soil (g/m²-yr), and AT is the accumulation rate (yr).

in sediment - water

Mercury concentrations in sediment - water were calculated from the following :

a) Mercury concentrations in sediment -

4. Determination of mercury concentrations

water due to runoff ($C_{sed}^{(10)}$)

$$C_{sed} = \frac{(\chi_e)(WA_L)(C_{soil})(AT)}{(BD_{soil})(SD_{soil})(WA_w)} \quad \text{--- (13)}$$

where the WA_L is watershed area contributing to runoff (m²), AT is the accumulation time (yr), and WA_w is the area of waterbody receiving runoff (m²).

b) Mercury concentrations in sediment - water due to deposition ($C_{sed(dep)}^{(10)}$)

$$C_{sed(dep)} = \frac{(\chi_e)(AT)(TD)}{(BD_{soil})(SD_{soil})} \quad \text{--- (14)}$$

RESULTS

Figure 2 and 3 show the patterns and concentrations of mercury along the plume centerline for X - Z (100m x 50m scale), and X - Y (100m x 10m scale), respectively. The concentrations were calculated in every cell along both X - Z and X - Y axis. These calculations (ng/m^3) were examples of the concentrations under neutral stability (D - stability) and u_s of 2 m/s. The plume spread slowly under neutral stability, thus GLC further away (downwind) from the incinerator at 3.7 to 5 km were the highest (approximately $28 \text{ ng}/\text{m}^3$).

WR was calculated from data of mercury concentrations in the air and rain collected by Greengerg *et al.* (1992)⁽¹⁴⁾. Estimated WR was 5.04×10^3 . Then, WD was approximately $4.2 \times 10^2 \text{ g}/\text{m}^2/\text{yr}$. The estimated DD using v_a of $0.006 \text{ cm}/\text{s}$ was $5.29 \times 10^7 \text{ g}/\text{m}^2\text{-yr}$. Thus, TD of atmospheric mercury to surfaces in the area was $4.73 \times 10^4 \text{ g}/\text{m}^2/\text{yr}$.

Estimated LS was 3.39 that was used to estimate χ_c of $27.8 \text{ g}/\text{m}^2\text{-yr}$. BD_{soil} is 0.01 m for standard assumption and deposited mercury is obviously retained at the soil surface,^(15, 16) SD_{soil} is $1300 \text{ Km}/\text{m}^2$ for the most likely value⁽¹⁰⁾. Results showed the K_1 was $2.14 \times 10^{-5} \text{ yr}^{-1}$ and C_{soil} was $0.37 \text{ mg}/\text{kg}$. K_{SE} was $0.35 \text{ ton}/\text{acre}$ (assuming loam or silty loam).⁽¹³⁾ The c value of 0.003, the most likely value, was used because the area around the incinerator was assumed to consist of 20 to 70% undisturbed land with greater than 40% litter cover. The default for P is 0.5⁽¹³⁾. The D value of 0.4 was given by Bonazountas and Wagner (1984)⁽¹⁷⁾. Estimated C_{sed} was $0.27 \text{ mg}/\text{kg}$ and $C_{sed/water}$ was $0.063 \text{ mg}/\text{kg}$ when WA_l of $10,000 \text{ m}^2$ and WA_w of $50,000 \text{ m}^2$ were used (derived from the GIS database of the surrounding watershed)⁽¹⁰⁾. Thus, total mercury in sediment - water was approximately $0.33 \text{ mg}/\text{kg}$.

DISCUSSIONS AND CONCLUSIONS

Estimated (high) mercury concentrations of $28 \text{ ng}/\text{m}^3$ (regardless background concentrations) were higher than average in the US ($2 - 20 \text{ ng}/\text{m}^3$)⁽⁸⁾ due to the near - source calculations and assumptions used in the calculations (no chemical reactions (loss), total mercury estimations). Background concentrations were reported by NJDEPE (1993) to be about $2 - 5 \text{ ng}/\text{m}^3$. WD of mercury accounted for 89% of TD because the wet deposition processes are an important removal process of soluble - mercury species (such as HgCl) near the incinerator. In addition, C_{rain} and C_{at} data from Greenberg *et al.* (1992) provided WR of 5.04×10^3 which is higher than data reported value

(not in New Jersey) by Brosset (1987)⁽¹⁸⁾ of 1×10^4 to 5×10^4 because the WR was derived from near - source measurement (2 to 3 km). From the results of the depositions, mercury in soil, and sediment - water of 0.37 to $0.33 \text{ mg}/\text{kg}$ were also high comparing to another county (Atlantic County) soil samples (<0.062 to $<0.089 \text{ mg}/\text{kg}$)^(1, 3). Results from the calculations indicated that high levels of mercury can be found in the air, soil, and sediment - water in the vicinity of the incinerator. Thus, the incinerator can be considered a major source of mercury pollution in the area in accordant with the reports by Aucott and Winka (1996), and NJDEPE (1993). Further information about

mercury in biota, e.g., fish, can be calculated using bioconcentration factor available in literatures.

Application of Gaussian dispersion equation shows that it can be a very useful tool for preliminary estimation of mercury dispersion point sources for certain assumptions. More accurate calculations can

be accomplished with complex models that account for chemical reactions of mercury or complex terrain.

Universal loss equation, depositions, and other accumulation equations are some of the methods used by regulatory agencies to estimate the concentrations of mercury in the environment

REFERENCES

1. Lee, Y - H, and Iverfeldt, A. (1991) Measurement of methylmercury and mercury in run - off, lake and rain waters. *Water, Air, and Soil Pollution*, 56 : 309 - 321.
2. Suzuke, T., Imura, N., and Clarkson, T.M. (1991). *Advance in mercury toxicology*. Plenum Press, NY.
3. Massachusetts Department of Environmental Protection (MADEP) (1996). *Mercury in Massachusetts : An evaluation of sources, emission, impacts and controls* Office of Research and Standards.
4. Aucott, M., Winka, M. (1996) Findings and recommendations of the New Jersey Mercury emissions standard setting task force. *J. of Hazardous Material*, 47 : 103 - 117.
5. Cooper, C.D., and Alley, F.C. (1994). *Air Pollution and Design Approach*, Second Ed., Waveland Press, Inc., IL.
6. Turner, B.D. (1970). *Workbook of atmospheric dispersion estimates*. Environmental Protection Agency, Office of Air Programs, NC.
7. U.S. Environmental protection Agency (USEPA) (1995). *User's guide for the industrial source complex (ISC3). Volume II, Description of model algorithms*. Office of Air Quality Planning and Standards Emissions, Monitoring, and / Analysis Division, NC.
8. Davis, M.L., and Cornwell, D.A. (1991). *Introduction to Environmental Engineering*, Second Ed., McGraw - Hill, Inc., NY
9. Iverfeldt, A. (1991). Mercury in forest canopy throughfall water and its relation to atmospheric deposition. *Water, Air, and Soil Pollution*, 56 : 553 - 564.
10. New Jersey Department of Environmental Protection and Energy (NJDEPE, 1993), Task Force on Mercury Emissions Standard Setting. B. II, Trenton, NJ.
11. Shannon, J.D., and Voldner, E.C. (1995). Modeling atmospheric concentrations of mercury and deposition to the Great Lakes. *Atmospheric Environment*, 29 : 1649 - 1661.
12. Hicks, B.B., Baldocchi, D.D., Meyers, T.P., Hosker, R.P., Jr., Matt, D.R. (1987). A preliminary multiple resistance routine for deriving deposition velocities from measured quantities, *Water, Air, And Soil Pollution*, 36 : 311-330.

13. Arnold, J.G., Williams, J r , Nicks, A.D., and Sammons, N.B. (1990) *SWRRB · A Basin - Scale simulation model for soil and water resources management*. Texas A&M University Presss, TX (as cited by NJDEPEE, 1993 a)
14. Greenberg, A., Wojtenko I., Chen, H.W , Krivanek, S., Butler, J P , Held, J , Weis P , and Reiss. N.M. (1992). *Mercury in air and rainwater in the vicinity of a municipal resource recovery facility in northwester New Jersey* Paper Presented in AWMA MEETING 1992, DURHAM, NC
15. Aastrup, A , Johnson, J., Brignmark, E , and Iverfeldt, A. (1991) Occurrence and transport of mercury within a small catchment area. *Water, Air, and Soil Pollution*, 56 155 - 168.
16. Lodenus, M (1994) Mercury in terrestrial ecosystems A review. In *Mercury Pollution . Integration and synthesis*. pp 343 - 354 Watras, C.J and Huckabee, J W Editors, Lewis Publishers, FL
17. Bonazountas, M., and Wagner. J M. (1984). *SESOIL, A Seasonal Soil Compartment Model*. May (as cited by NJDEPE, 1993)
18. Stein, E D , Cohen, Y , and Winer, A.M. (1996). Environmental distribution and transformation of mercury compounds. *Critical Reviews in Env. Sci and Tech*. 26 1 - 43
19. Brosset, C. (1987). The behavior of mercury in the physical environent. *Water, Air, and Soil Pollution*, 34 : 145 - 166.
20. Radian Corporation (1989) *Municipal solid waste combustion Assessment of health risks associated with municipal waste combustion emission*, Hemisphere Publishing Corp., NY
21. Martin, J.R. (1988) *Obtaining air quality permits for a Resource Recovery Facility : A case study in Warren county, New Jersey*. Paper presented at 81st Annual Meeting of APCA, Texas
22. Chun, M. (1996). *Modeling of mercury wet deposition from an incinerator* Ph.D. Dissertation, Rutgers University, NJ.
23. Lindberg, S E., Meyers, T P , Taylor, G E Jr , Turner, R.R., and Schroeder, W H. (1992). Atmosphere - surface exchange of mercury in a forest : Results of modeling and gradient approaches. *J Geophys. Resear* 97 2519 - 2528

APPENDIX

Table 1. : Parameters used in calculations

| Parameters | Values | Remarks |
|-------------|---------------------------|--|
| Q | 0.013 $\mu\text{g/s}$ | Permit Value (NJDEPE, 1993) |
| h | 76.2 m | Radian Corporation (1989) |
| v_s | 18.29 m/s | Martin (1988) |
| d | 2.48 m | Radian Corporation (1989) |
| R | 0.9 | Reflection of plume 90% (assume 10% soluble) |
| P | 9.87×10^{-3} kPa | Standard pressure (1 atm) |
| T_s | 388.6°K | Martin (1988) |
| T_a | 293°K | Standard temperature (20°C) |
| u_s | 2 m/s | Chu (1996) |
| WR | 5.04×10^3 | Greenberg et al., 1992 |
| R_f | 1000 hr/yr | Chu (1996) NJ data |
| t_r | 0.003 m/hr | Chu (1996) NJ data |
| v_d | 0.006 cm/s | Lindberg et al., 1992 |
| S | 0.0433 m/m | Most likely value (NJDEPE, 1993) |
| | 2000 m | NJDEPT, (1993) |
| E_R | 175 yr^{-1} | NJDEPE, (1993) |
| K_{SE} | 0.35 ton/acre | NJDEPE, (1993) |
| c | 0.003 | Most likely value (NJDEPE, 1993) |
| P | 0.5 | NJDEPE, (1993) |
| D | 0.4 | Bonazauntas and Wagner, 1984 |
| BD_{soil} | 0.01 m | Standard value (NJDEPE, 1993) |
| SD_{soil} | 1300 kg/m^3 | Most likely value (NJDEPE, 1993) |
| AT | 10 yr | NJDEPE, (1993) |
| WA_L | $100,000 \text{ m}^2$ | NJDEPE, (1993) |
| WA_W | $50,000 \text{ m}^2$ | NJDEPE, (1993) |

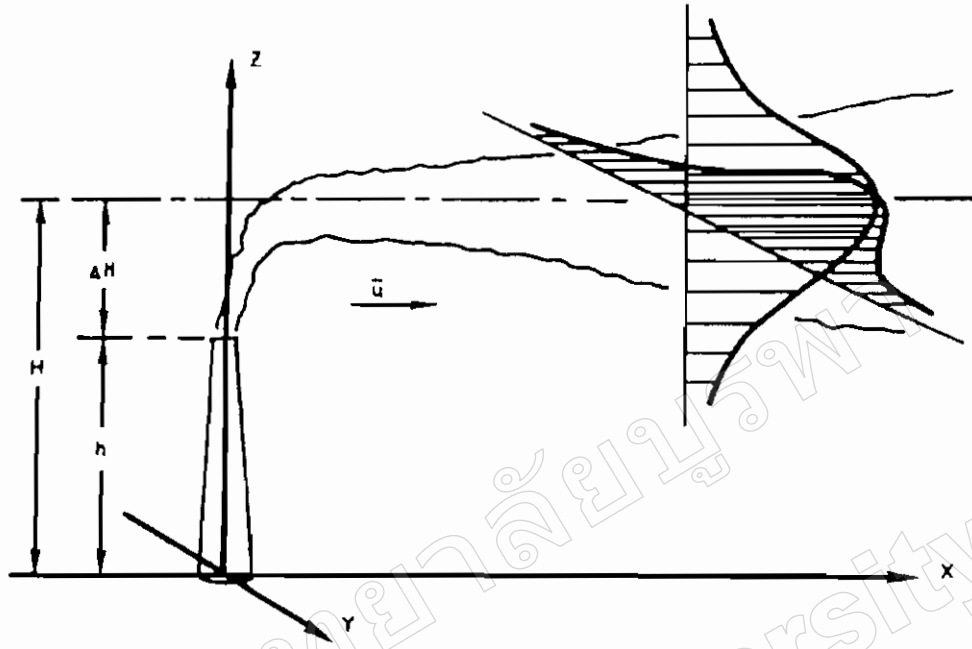


Figure 1 : Gaussian Plume Model

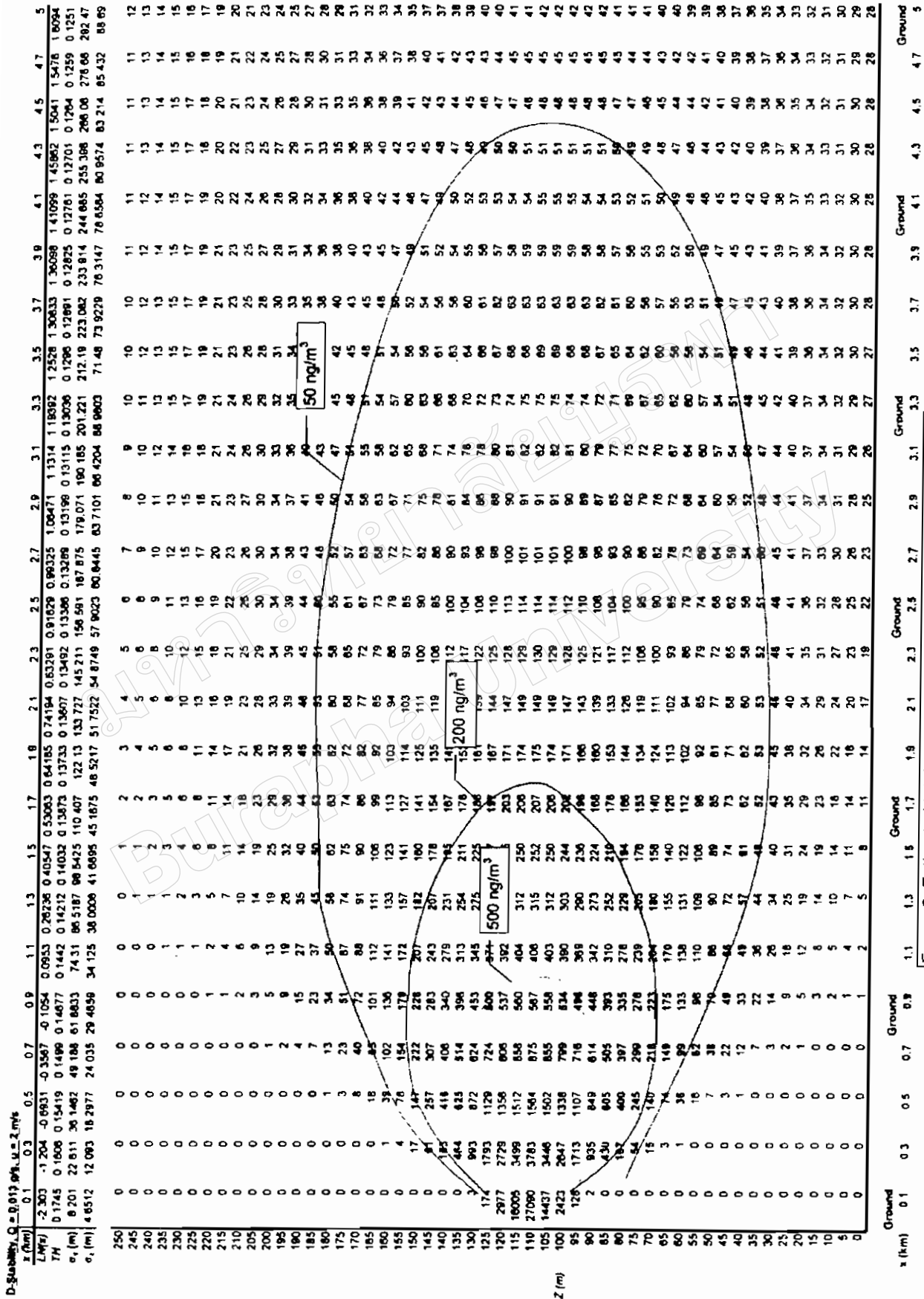


Figure 2: Estimated total mercury concentrations along x-z axis

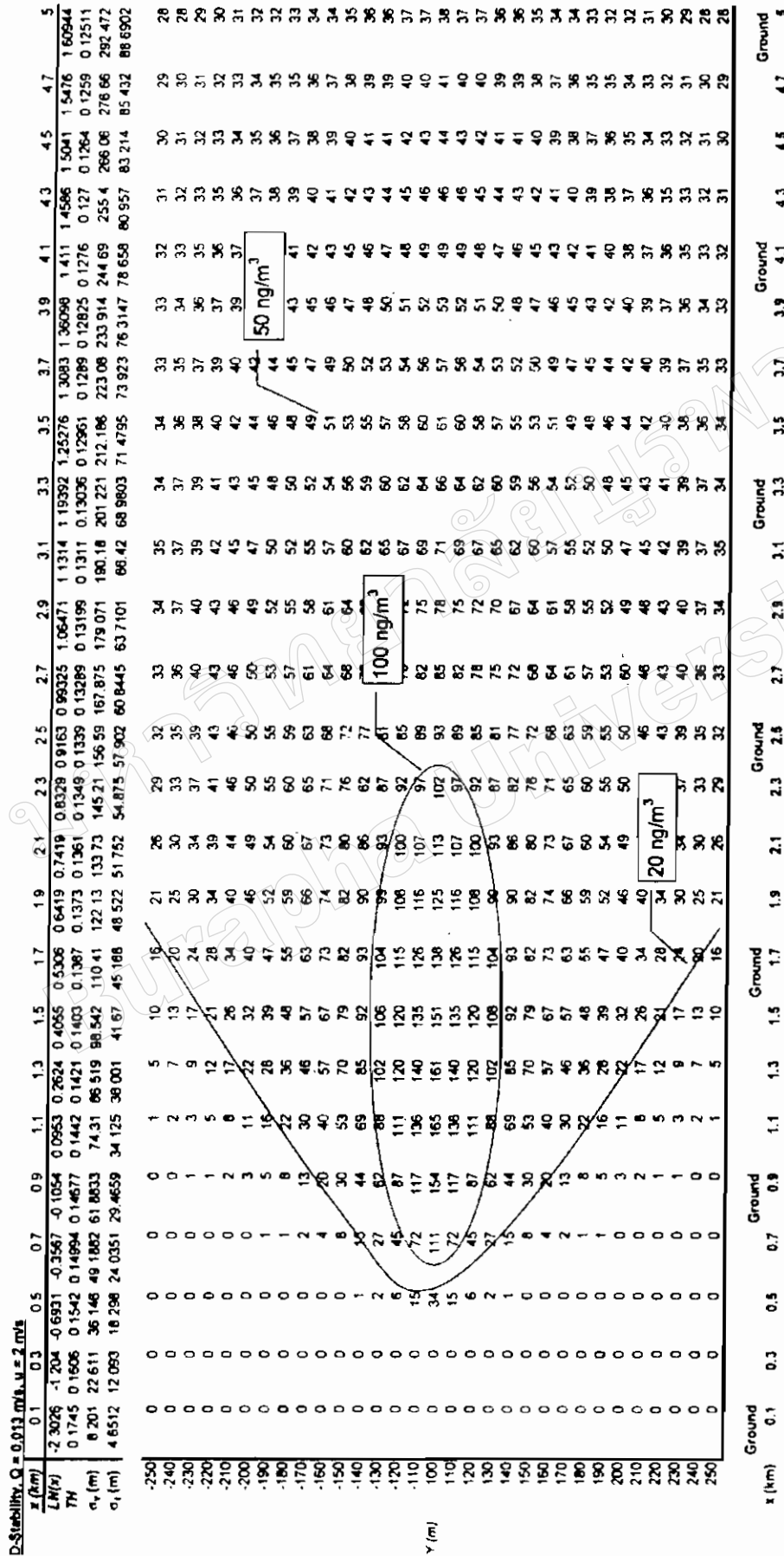


Figure 3: Estimated total mercury concentrations along x-y axis